

Removal of Pb(II) Ion from Aqueous Solution Using As-synthesized CuO/Cu₂O Nanocomposites

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In this paper, the spherical CuO/Cu₂O nanocomposites were synthesized using co-precipitation accompanied by annealing at 500 and 600 °C. The as-synthesized CuO/Cu₂O nanocomposites were characterized by Fourier transform infrared (FT-IR) spectroscopy, X-ray diffraction (XRD) and transmission electron microscope (TEM). The XRD and FT-IR results proved the successful synthesis of the CuO/Cu₂O nanocomposite. Spherical shapes of the samples confirmed by the TEM images with narrow particle size distribution. The average size of the nanocomposites synthesized at 600 °C (39 nm) was smaller than that of the nanocomposites synthesized at 500 °C (46 nm). In addition, the samples were chemically activated using H₂O₂ and used as new adsorbents to remove the Pb(II) ion from an aqueous solution. The effect of solution pH, sorbent dose, initial Pb(II) concentration, and the contact time were studied. Results showed that the highest efficiency (85% for the nanocomposites synthesized at 500 °C and 92% for the nanocomposites synthesized at 600 °C) was obtained at pH 6, 90 min contact time, 30 ppm Pb(II) solution, and 0.02 g of the sorbent. The Pb(II) adsorption equilibrium data were fitted well to the Langmuir model.

Keywords: CuO/Cu₂O nanocomposites, Co-precipitation, Adsorbent, Pb(II) Removal, Langmuir

INTRODUCTION

Nowadays, water pollution caused by toxic and dangerous organic dyes [1,2] and heavy metal ions [3,4] is one of the most important challenges all around the world. The presence of toxic heavy metal ions in water has adverse effects on animals, plants, and even humans. They are stable in the nature and do not disappear easily and quickly. Therefore, it is necessary to remove these ions from wastewater before discharging into environment; this is one

of the most important goals of chemical engineering [5-8]. To do so, several methods have been reported such as chemical precipitation, ion-exchange, adsorption, membrane filtration, coagulation–flocculation, flotation, and electrochemical [9,10]. Among these methods, the adsorption process is the most popular one due to its simplicity and high efficiency. Also, it does not require complex and expensive equipment [7,8,11,12]. However, the choice of adsorbent type plays an important role in this method [7,8,11,12]. Recently, the application of transition metal oxide nanoparticles for the adsorption and removal of heavy metal ions, due to their large surface area and availability, has been an interesting area of research [3,5,6].

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Among them, the copper oxides (CuO and Cu₂O) have attracted attention due to their properties (narrow band gap as p-type semiconductors) and their practical applications [13-15]. In recent years, many articles have been reported on the use of copper oxide nanoparticles for removal of metal ions [16,17]. For example, Raul *et al.* used copper oxide nanorods to remove Pb(II) ions from an aqueous solution. The study revealed that 92% of Pb(II) ions was removed in 60 min using this adsorbent [18]. Din Mahmoud *et al.* [16] used copper oxide nanoparticles synthesized by green chemistry to remove Pb(II), Ni(II) and Cd(II) ions. Their results showed that the removal efficiency of Pb(II) ions was higher than that of Ni(II) and Cd(II) ions.

Continuing the previous work on the preparation of various transition metal nanoparticles [19-21], the main purpose of this paper is to prepare CuO/Cu₂O nanocomposites and use them as efficient adsorbents to remove Pb(II) ions from an aqueous solution.

EXPERIMENTAL

Materials and Methods

Copper(II) acetate mono hydrate, oxalic acid, and lead(II) nitrate were purchased from Merck Company and used without further purification. FT-IR was recorded using a Perkin-Elmer spectrophotometer from 4000 to 400 cm⁻¹. The XRD patterns were obtained on Empyrean powder diffractometer of PANalytical in the range 2θ = 10-80° (Cu Kα radiation, λ = 1.5418 Å). TEM images were obtained by transmission electron microscope Philips CM120 with LaB₆ cathode operating at 120 kV and equipped with CCD camera Olympus Veleta.

Preparation of CuO/Cu₂O Nanocomposites

3 g of Cu(CH₃COO)₂·H₂O and 3 g oxalic acid were poured into a beaker containing 20 ml of water and stirred for 3 h. The final mixture was completely dried at 90 °C for 6 h. After that, the resulting precipitate was ground in a crucible and annealed in the electric furnace at 500 and 600 °C for 3 h. The resulting black precipitates were washed with distilled water, filtered, dried under ambient atmosphere, and characterized by FT-IR, XRD and TEM.

Removal of Pb(II) Ion

Solutions with concentrations of 10, 30, 50, 70, 90, and

110 ppm were prepared by dissolving lead nitrate in distilled water. In order to predict the adsorption behavior, a series of experiments were conducted as a function of contact time, initial Pb(II) concentration, pH solution, and sorbent dose. Sample in a 100 ml Erlenmeyer flask containing 50 ml of the Pb(II) solution were continuously stirred in an air bath shaker at 200 rpm and constant temperature of 30 for certain times varying from 0 to 120 min. In order to produce an activated sorbents, about 2 ml of H₂O₂ was added to the suspension. The solution pH was adjusted as needed using HCl (0.1 M) or NaOH (0.1 M). Then, at different stirring times, after removing adsorbent by centrifuge, the concentration of Pb(II) ions was measured using atomic adsorption. All the experiments were replicated three times. The Pb(II) removal efficiency (%) and adsorption capacity (mg g⁻¹) were calculated using the following formulas:

$$R (\%) = \{(C_0 - C_t)/C_0\} \times 100 \quad (1)$$

$$q (\text{mg g}^{-1}) = \{(C_0 - C_t)/M\} \times V \quad (2)$$

Here, C₀ and C_t are the initial and final concentrations of Pb(II) ions in solution, respectively. V and M are the volume of the solution and the mass of the adsorbent, respectively.

The adsorption kinetic for the adsorption of Pb(II) ion onto as-prepared CuO/Cu₂O nanocomposites was studied by the two different models of pseudo-first and second-order models, and the equations are given below:

$$\ln(q_e - q_t) = \ln q_e - k_1 t \quad (3)$$

$$t/q_t = 1/k_2 q_e^2 + t/q_e \quad (4)$$

Here, k₁ (min⁻¹) and k₂ (g mg⁻¹ min⁻¹) are the pseudo-first and second-order rate constant values. q_e and q_t (mg g⁻¹) are the adsorption capacity of as-prepared CuO/Cu₂O nanocomposites at equilibrium and at time t (min), respectively.

Lanmuir's and Freundlich isotherm models were generally studied to sorption mechanism of Pb(II) metal ions as-prepared CuO/Cu₂O nanocomposites and the equations are given below:

$$C_e/q_e = 1/q_t b + C_e/q_t \quad (5)$$

$$\ln q_e = \ln k_F + \ln C_e/n \quad (6)$$

Here, b (lmg^{-1}) and K_F are the Langmuir and Freundlich constants.

RESULTS AND DISCUSSION

Characterization of CuO/Cu₂O Nanocomposites

FT-IR, XRD and TEM techniques were used to characterize synthetic CuO/Cu₂O nanocomposites. Figure 1 shows the FT-IR spectra of the copper precursor and of the as-synthesized CuO/Cu₂O nanocomposites. In the FT-IR spectrum of the copper precursor, broad peaks, observed at 3423 and 1642 cm^{-1} , belong to the hydroxyl and carbonyl groups of the oxalic acid, respectively. The sharp peak at 489 cm^{-1} confirms the presence of Cu-O bonds in the compound. By thermal decomposition of the copper precursor at temperatures of 500 and 600 °C, the peaks of the oxalic acid group were removed. In the FT-IR spectrum of the as-synthesized CuO/Cu₂O nanocomposites, several peaks were obtained in the range of 480 to 620 cm^{-1} that are attributed to the vibrations of Cu^I-O and Cu^{II}-O bonds [21]. In these compounds, a weak and broad peak at 1620 cm^{-1} can be assigned to the moisture or water molecules adsorbed on the nanoparticles [21].

The XRD patterns of the as-synthesized CuO/Cu₂O nanocomposites are shown in Fig. 2. The peaks that appeared at the angles of 29.55°, 36.40°, 42.28°, 61.33°, and 73.46° can be assigned to the (110), (111), (200), (220), and (311) crystal planes of the cubic Cu₂O (ICSD 172174) [13-15]. The peaks appearing at the angles of 32.49°, 35.40°, 35.54°, 38.66°, 38.90°, 48.77°, 53.43°, 58.16°, 61.52°, 65.74°, 66.23°, 68.04°, 72.27°, 74.91°, and 75.24° correspond to the crystal planes (110), (002), (11-1), (111), (200), (20-2), (020), (202), (11-3), (022), (31-1), (220), (311), (004), and (22-2) of the monoclinic CuO (ICSD 291390) [13-15]. Moreover, there is one tiny impurity peak at 43.35° in the sample synthesized at 500 °C, which disappears at higher temperature. Otherwise, the sharpness and broadness of the peaks confirm good crystallinity and purity of the as-synthesized CuO/Cu₂O nanocomposites.

Rietveld fits revealed that CuO/Cu₂O nanocomposite

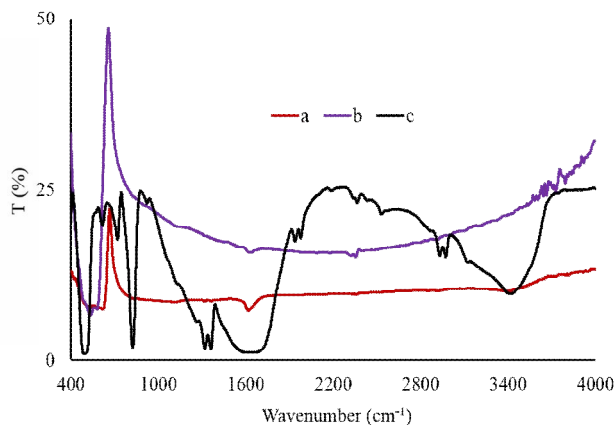


Fig. 1. FT-IR spectra of CuO/Cu₂O nanocomposites prepared at a) 500 °C, b) 600 °C and c) Cu precursor.

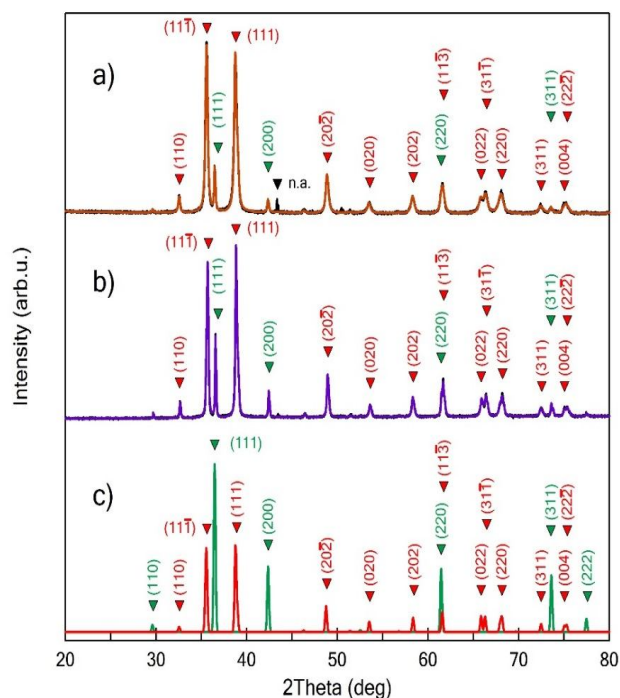


Fig. 2. XRD patterns of CuO/Cu₂O nanocomposites prepared at a) 500 °C and at b) 600 °C. Black dots represent measured points; color lines depict corresponding Rietveld fits. Green pattern in c) shows diffraction peaks of cubic Cu₂O (ICSD 172174), the red one diffraction peaks of monoclinic CuO (ICSD 291390).

Table 1. Structural Parameters of CuO/Cu₂O Nanocomposites Synthesized at 500 and 600 °C

		Temperature (°C)	
		500	600
Cu ₂ O (Pn-3m)	Fraction (%)	7.5	14.8
	a (Å)	4.2691(2)	4.2715(1)
CuO (C2/c)	Fraction (%)	92.5	85.2
	a (Å)	4.6851(2)	4.6887(1)
	b (Å)	3.4280(2)	3.4264(1)
	c (Å)	5.1313(2)	5.1346(1)
		β (°)	99.378(2) 99.345(1)

synthesized at 500 °C consisted of 7.5 % cubic Cu₂O (space group ‘Pn-3m’) and 92.5 % monoclinic CuO phase (space group ‘C2/c’); at higher temperature, the content of cubic Cu₂O raised to 14.8 %. All structural parameters for the both samples are summarized in Table 1.

It is known that line broadening of diffraction peaks in the XRD patterns of samples is influenced by the average crystallite size [22]. The average crystallite size of as-synthesized nanocomposites can be calculated using Scherrer’s equation [22]. It was about 39 nm for sample

synthesized at 600 °C and about 46 nm for sample synthesized at 500 °C.

$$D = 0.94\lambda/\beta\cos\theta \quad (7)$$

Here, λ is the wavelength (0.154, CuK α), β is the full width at half maximum (FWHM), and θ is the diffraction angle. The two sharp peaks appeared at 2θ values of 35.6 and 38.8° were chosen to calculate the crystallite size of the as-synthesized nanocomposites.

The shape and the size of the as-prepared CuO/Cu₂O nanocomposites were examined by the TEM. Figure 3 shows the obtained TEM images. Despite the effort to separate individual particles in ultrasound, there were particle clusters in the figures. The particles had almost spherical or elliptical shapes. The size of the particles was in the ranges of several nanometers to low tens of nanometers. The average equivalent diameter was estimated to be approximately 20 nm. Accurate image analysis of the particle size distribution was not possible due to the enormous particle overlap in the clusters.

Pb(II) Removal Study

Important parameters such as pH, shaking time, adsorbent dose, and the initial concentration of heavy metal

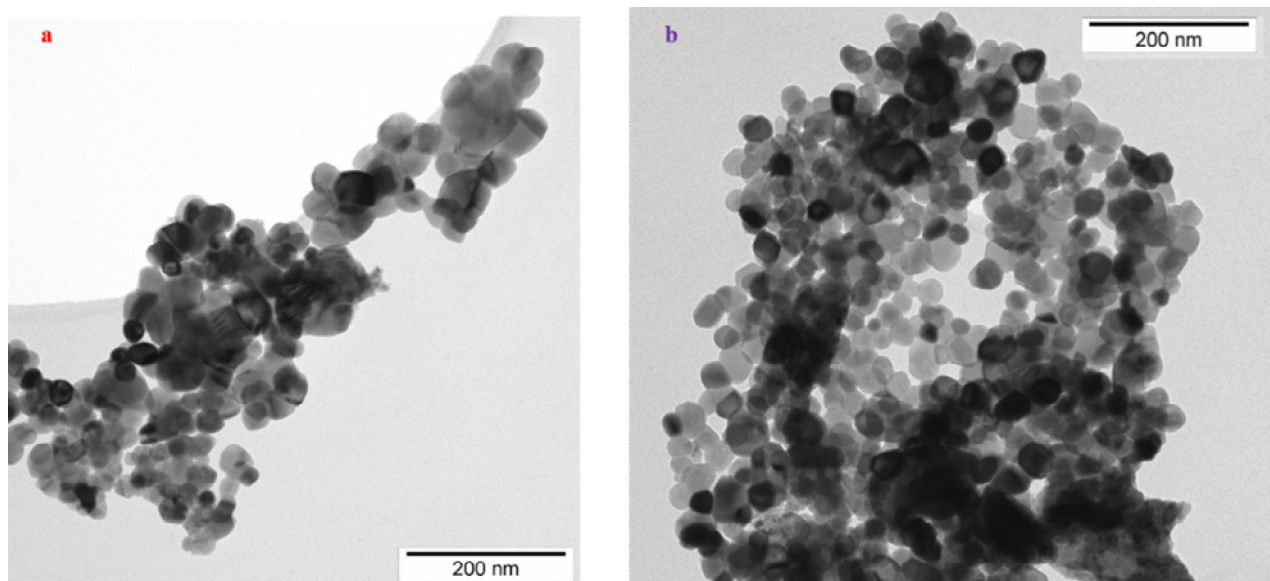


Fig. 3. TEM images of CuO/Cu₂O nanocomposites prepared at a) 500 °C and b) 600 °C.

ions can influence the removal percentage of heavy metal ions using various adsorbents [7-12,16,17]. At pHs < 4, the active groups on the surface of the adsorbent were protonated and could not be coordinated to the Pb(II) ions. Also, an electrostatic repulsion between the Pb(II) lead ions and the protonated groups on the surface of adsorbent caused low removal efficiency. However, at pHs > 7, the removal efficiency the Pb(II) ion was also low due to the precipitations of $\text{Pb}(\text{OH})^+$ and $\text{Pb}(\text{OH})_2$ [7-12,16,17]. Figure 4 shows the effect of pH on the removal of Pb(II) ions. It was observed that the maximum adsorptive removal of Pb(II) occurred at pH 6 [16,17].

The adsorbent dose is another important parameter [7-12]. Figure 5 shows the effect of adsorbent dose on the adsorption of Pb(II) ion using as-prepared CuO/Cu₂O nanocomposites at room temperature. It was obtained that the removal of Pb(II) was increased by an increases in adsorbent dose. This observance was due to an increase in active groups from 0.005 to 0.02 g; the removal percentage reached up to 85% for the CuO/Cu₂O prepared at 500 °C and 92% for the CuO/Cu₂O prepared at 600 °C. At low adsorbents dose, the active groups are fully coordinated by the Pb(II) ions. Therefore, large amount of Pb(II) ions cannot be coordinated to sorbent and remains in the solution [16,17]. The subsequent increase of the adsorbent dose does not cause any noticeable change in the absorption efficiency.

Figure 6 shows the effect of shaking time on the adsorption of Pb(II) ion. It was shown that in the first 15 min, the amount of Pb(II) ion removal was very intense. This is due to the high number of active groups in adsorbent and also the high concentration of Pb(II) ions. After that, due to the decrease in the amount of Pb(II) ions and also the active groups on the adsorbents, the removal of Pb(II) ions continued at a slower rate and reached its maximum value in 90 min [16,17].

Figure 7 shows the effect of initial concentration of Pb(II) ion on its removal efficiency by as-synthesized CuO/Cu₂O nanocomposites. The removal efficiency by the adsorbent decreased by an increase in the initial concentration of Pb(II) ions from 10 to 110 ppm. At low Pb(II) concentration (10 ppm), the better interaction of active groups on the adsorbent surface with Pb(II) ions and also the rapid and easy migration of the Pb(II) ion to the

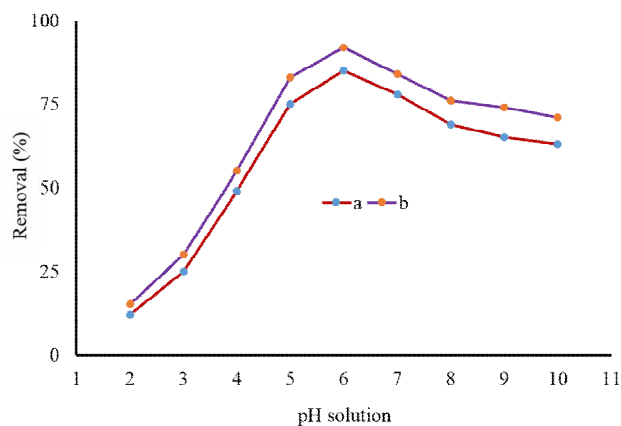


Fig. 4. Effect of pH on the Pb(II) removal using CuO/Cu₂O nanocomposites (Pb(II) concentration: 30 ppm, Shaking time: 90 min and adsorbent dose: 0.02 g).

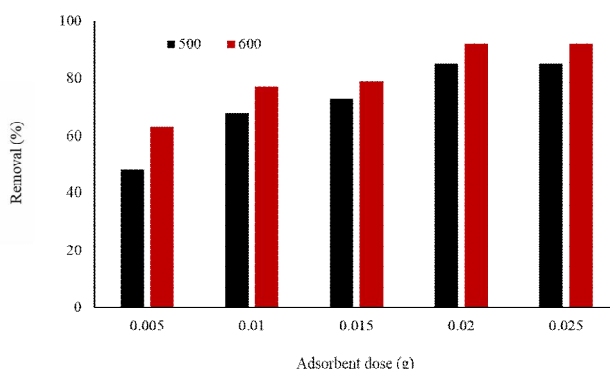


Fig. 5. Effect of sorbent dose on the Pb(II) removal using CuO/Cu₂O nanocomposites (Pb(II) concentration: 30 ppm, Shaking time: 90 min and pH solution: 6).

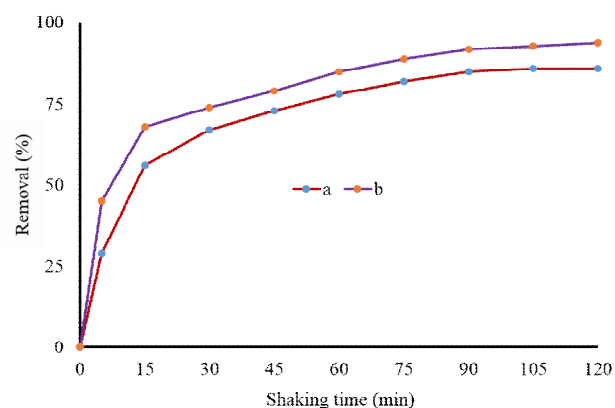


Fig. 6. Effect of shaking time on the Pb(II) removal using CuO/Cu₂O nanocomposites (Pb(II) concentration: 30 ppm, adsorbent dose: 0.02 g and pH solution: 6).

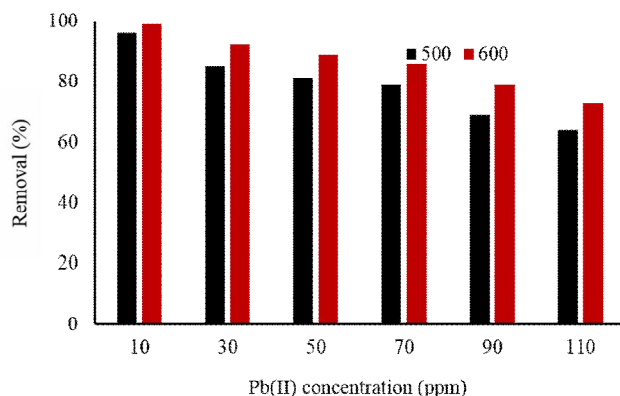


Fig. 7. Effect of initial Pb(II) concentration on the Pb(II) removal using CuO/Cu₂O nanocomposites (shaking time: 90 min, adsorbent dose: 0.02 g and pH solution: 6).

adsorbent surface resulted in a high efficient removal of the Pb(II) ion (96% for the sample prepared at 500 °C and 99% for the sample prepared at 600 °C) from aqueous solution. An increase in the concentration of the Pb(II) ions prevented some of the Pb(II) ions from interacting with the active site

of sorbent (due to occupation of them by Pb(II) ion) [23-27]. In this case, some of the Pb(II) ions could not be removed from the environment [26,17]. Adsorption capacities of samples were calculated and they are shown in Table 2.

Table 3 represents the comparison of the adsorption capacity of other sorbents and as-prepared current CuO/Cu₂O nanocomposites for the Pb(II) ion. Results revealed that the as-prepared CuO/Cu₂O nanocomposites was a promising adsorbents for the removal of Pb(II) ion.

In the real water, there are different cations (Na⁺, Ca²⁺, Mg²⁺) and anions (NO₃⁻, Cl⁻, PO₄³⁻) that can affect the adsorption performance of the adsorbents [31-33]. The effect of these coexisting ions on the Pb(II) removal using as-synthesized CuO/Cu₂O nanocomposites were studied.

Adsorption Kinetics

Figures 8 and 9 represent the pseudo-first and second-order adsorption kinetics of Pb(II) ion onto the as-synthesized CuO/Cu₂O nanocomposites, respectively. By comparing the straight lines of the adsorption profiles, it can be seen that adsorption kinetics followed the pseudo

Table 2. Adsorption Capacities of As-prepared CuO/Cu₂O Nanocomposites

Pb(II) concentration		10	30	50	70	90	110
q (mg g ⁻¹)	CuO/Cu ₂ O Prepared at 500 °C	24.0	63.7	105.0	138.2	155.2	176
	CuO/Cu ₂ O Prepared at 600 °C	24.7	69.0	111.2	150.5	177.7	200.7

Table 3. A comparative of Adsorption Capacities for Pb(II) Using Various Adsorbents

Adsorbents	Adsorption capacity (mg g ⁻¹)	Ref.
CuO/Cu ₂ O	63.7 and 69	This work
CuO	37.02	[17]
CuO	88.8	[16]
Cu/Mg Binary ferrite	57.7	[23]
ECCSB@Fe ₃ O ₄	86.2	[11]
m-CSPIB	104.16	[12]
NPCS-PEI	645.16	[28]
Sulphydryl modified chitosan	273.7	[29]
Chitosan/lignosulfonate	525	[30]

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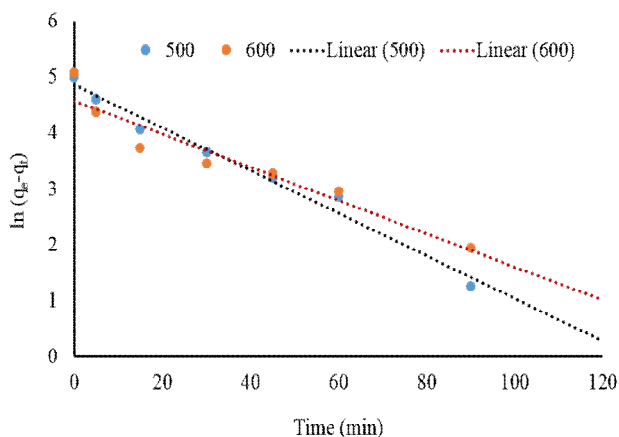


Fig. 8. Pseudo first-order equation curve for Pb(II) ion adsorption onto CuO/Cu₂O nanocomposites.

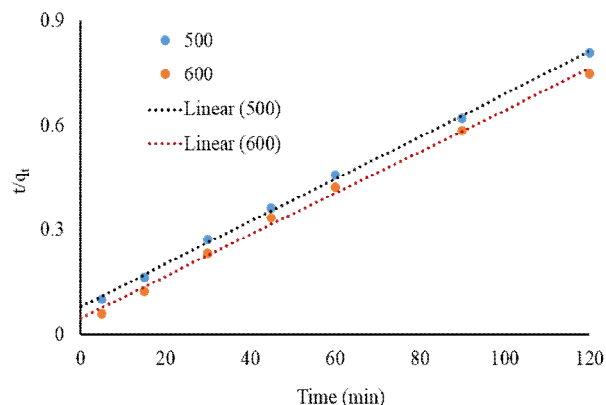


Fig. 9. Pseudo second-order equation curve for Pb(II) ion adsorption onto CuO/Cu₂O nanocomposites.

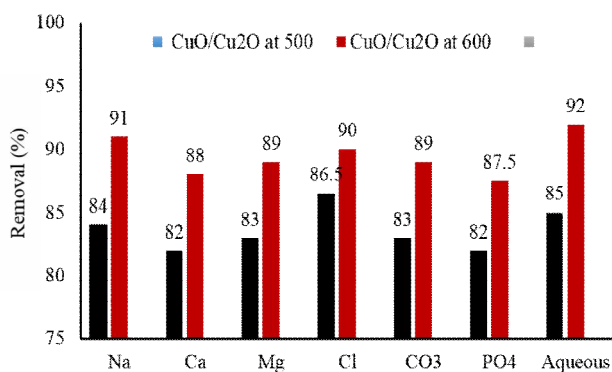


Fig. 8. The effect of coexisting ions on the Pb(II) removal using as-synthesized CuO/Cu₂O nanocomposites.

second-order model [17,23,27], which confirmed that the adsorption capacity was related to the active sites on the surface of the as-synthesized CuO/Cu₂O nanocomposites.

Adsorption Isotherms

In order to investigate mechanism of Pb(II) ion adsorption onto the as-synthesized CuO/Cu₂O nanocomposites, the Langmuir and Freundlich isotherm were studied and the curves are given in Figs. 10 and 11, respectively. It was observed that the Freundlich isotherm was more suitable for adsorption of Pb(II) ion on the surface of the as-synthesized CuO/Cu₂O nanocomposites [34]. This indicates that the adsorption of Pb(II) ion was carried out by multiple and heterogeneous layers of the surface of the samples [35].

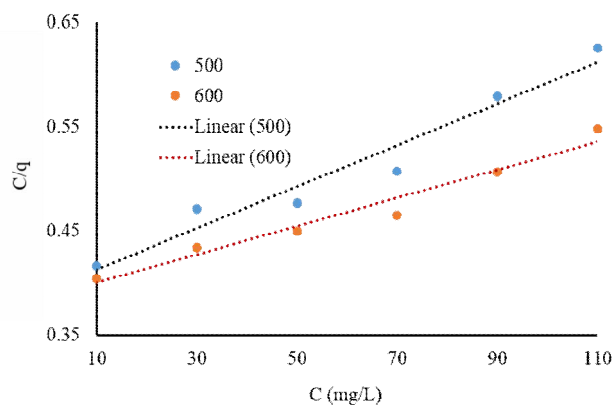


Fig. 10. Langmuir isotherm for Pb(II) ion adsorption onto CuO/Cu₂O nanocomposites.

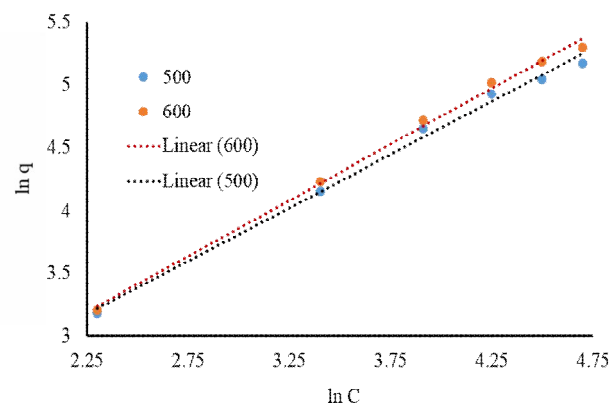


Fig. 11. Freundlich isotherm for Pb(II) ion adsorption onto CuO/Cu₂O nanocomposites

CONCLUSIONS

The pure CuO/Cu₂O nanocomposites were successfully synthesized using thermal decomposition at 500 and 600 °C as a facile and low-cost route. FT-IR and XRD patterns confirmed the preparation of pure and single crystalline phase of CuO/Cu₂O nanocomposites. The adsorption of Pb(II) using the as-prepared CuO/Cu₂O nanocomposites was found to be dependent on the pH solution, sorbent dose, shaking time, and the initial concentration of Pb(II). The maximum removal efficiencies of Pb(II) was 92% and 85% using samples at the optimum equilibrium contact time of 120 min, sorbent dose of 0.02 g, and initial Pb(II) concentration of 70 mg/L. The experimental data predicted that the Pb(II) adsorption followed the pseudo second-order, and the adsorption process was fitted to the Freundlich isotherm model.

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CONFLICT OF INTEREST

The authors declare that they have no conflict of interest.

AUTHOR CONTRIBUTIONS

M. Jarosova and P. Machek gathered XRD patterns and TEM images. A.D. Khalaji prepared the samples and recorded FT-IR spectra. S.A. Abed, R. Solanki, F.H. Alsultany, and S. Abed Hussein prepared samples and investigated the photocatalytic activity of samples. A.D. Khalaji wrote the main manuscript. All authors reviewed and revised the MS.

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